




Long-term exposure to PM₁ is associated with increased prevalence of metabolic diseases: evidence from a nationwide study in 123 Chinese cities

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Abstract

Exposure to particulate matter (PM) has been linked to metabolic diseases. However, the effects of PM with an aerodynamic diameter $\leq 1.0 \mu\text{m}$ (PM₁) on metabolic diseases remain unclear. This study is aimed at assessing the associations of PM₁ with metabolic disease risk and quantifying the concentration–response (C-R) relationship of PM₁ with metabolic disease risk. A national cross-sectional study was conducted, including 12,495 middle-aged and older adults in 123 Chinese cities. The two-year average concentration of PM₁ was evaluated using satellite-based spatiotemporal models. Metabolic diseases, including abdominal obesity, diabetes, hypertension, dyslipidemia, and metabolic syndrome, were identified based on physical examination, blood standard biochemistry examination, and self-reported disease histories. Generalized linear models and C-R curves were used to evaluate the associations of PM₁ with metabolic diseases. A total of 12,495 participants were included in this study, with a prevalence of 45.73% for abdominal obesity, 20.22% for diabetes, 42.46% for hypertension, 41.01% for dyslipidemia, and 33.78% for metabolic syndrome. The mean \pm standard deviation age of participants was 58.79 ± 13.14 years. In addition to dyslipidemia, exposure to PM₁ was associated with increased risks of abdominal obesity, diabetes, hypertension, and metabolic syndrome. Each $10 \mu\text{g}/\text{m}^3$ increase in PM₁ concentrations was associated with 39% (odds ratio (OR) = 1.39, 95% confidence interval (CI) 1.33, 1.46) increase in abdominal obesity, 18% (OR = 1.18, 95%CI 1.12, 1.25) increase in diabetes, 11% (OR = 1.11, 95%CI 1.06, 1.16) increase in hypertension, and 25% (OR = 1.25, 95%CI 1.19, 1.31) in metabolic syndrome, respectively. C-R curves showed that the OR values of abdominal obesity, diabetes, hypertension, and metabolic syndrome were increased gradually with the increase of PM₁ concentrations. Subgroup analysis indicated that exposure to PM₁ was associated with increased metabolic disease risks among participants with different lifestyles and found that solid fuel users were more susceptible to PM₁ than clean fuel users. This national cross-sectional study indicated that exposure to higher PM₁ might increase abdominal obesity, diabetes, hypertension, and metabolic syndrome risk, and solid fuel use might accelerate the adverse effects of PM₁ on metabolic syndrome risk. Further longitudinal cohort studies are warranted to establish a causal inference between PM₁ exposure and metabolic disease risk.

Keywords Particulate matter · Obesity · Diabetes · Hypertension · Dyslipidemia · Metabolic syndrome

Introduction

Non-communicable diseases have been reported as the leading causes of global burden, accounting for 15 million premature deaths each year (Chong et al. 2023; Collaborators 2020). A large percentage of non-communicable diseases could be attributed to the increasing burden of metabolic diseases, such as obesity, diabetes, hypertension, dyslipidemia, and metabolic syndrome (Chew et al. 2023; Chong et al. 2023; Han et al. 2022). Genetic defects, physical inactivity, unhealthy behaviors, and lifestyles have been identified

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as significant risk factors for rising metabolic disease risks (Chong et al. 2023). However, those factors might not fully explain the high prevalence of metabolic diseases, environmental risk factors, such as particulate matter (PM), have attracted extensive attention from epidemiologists in recent years.

Numerous studies indicated that exposure to PM was associated with increased metabolic disease risks, especially PM with an aerodynamic diameter $\leq 2.5 \mu\text{m}$ ($\text{PM}_{2.5}$) and PM with an aerodynamic diameter $\leq 10 \mu\text{m}$ (PM_{10}). Many studies indicated that exposure to PM was associated with an increased risk of metabolic diseases, including obesity (Zheng et al. 2021), diabetes (Liu et al. 2019a, 2019b; Yitshak Sade et al. 2023), hypertension (Li et al. 2020; Niu et al. 2023; Weng et al. 2022), dyslipidemia (Mao et al. 2020a), and metabolic syndrome (Han et al. 2022; Letellier et al. 2022; Zhang et al. 2021a). For example, Zheng et al. conducted a cross-sectional study of 36,456 aged 7–19 years in Jiangsu province of China and found that each $10 \mu\text{g}/\text{m}^3$ increase in $\text{PM}_{2.5}$ concentrations was associated with an 18.5% increase in obesity risk (odds ratio (OR) = 1.198, 95% confidence interval (CI) 1.054, 1.333) (Zheng et al. 2021). Liu et al. conducted a meta-analysis of 30 studies and indicated that long-term exposure to $\text{PM}_{2.5}$ and PM_{10} was all associated with increased diabetes. They found each $10 \mu\text{g}/\text{m}^3$ increase in $\text{PM}_{2.5}$ and PM_{10} concentrations was associated with 9% (OR = 1.09, 95%CI 1.05, 1.13) and 12% (OR = 1.12, 95%CI 1.06, 1.19) increase in diabetes risk, respectively (Liu et al. 2019a). Weng et al. conducted a prospective study of 91,366 participants in the UK biobank and found each $10 \mu\text{g}/\text{m}^3$ increase in $\text{PM}_{2.5}$ and PM_{10} was associated with 41% (hazard ratio (HR) = 1.41, 95%CI 1.29, 1.53) and 5% (HR = 1.05, 95%CI 1.00, 1.09) increase in hypertension incidence, respectively (Weng et al. 2022). Letellier et al. conducted cross-sectional study of 552 participants Hispanics/Latinos and non-Hispanics and indicated that higher dynamic exposure to $\text{PM}_{2.5}$ was associated with increased risk of obesity (RR = 1.23, 95% CI 1.10, 1.37), dyslipidemia (RR = 1.12, 95% CI 1.03, 1.22) and metabolic syndrome (RR = 1.17, 95% CI 1.07–1.28) (Letellier et al. 2022).

PM with an aerodynamic diameter $\leq 1.0 \mu\text{m}$ (PM_1), smaller size PM with a bigger surface-to-mass ratio than $\text{PM}_{2.5}$ and comprising over 80% of $\text{PM}_{2.5}$ (Li et al. 2019; Yang et al. 2018b), could carry more extensive toxins and easily penetrate deep into lung alveolar region and enter into the circulatory system (Guo et al. 2022; Yang et al. 2019). The health effects of PM_1 have drawn increasing attention from epidemiologists in recent years (Xu et al. 2022). However, compared with sufficient evidence of $\text{PM}_{2.5}$ and PM_{10} , the number of PM_1 is limited and results are still controversial. Several studies suggested that exposure to PM_1 was associated with increased metabolic disease risks (Hou et al.

2020; Li et al. 2019; Mao et al. 2020b; Yang et al. 2019; Yang et al. 2018a), while other studies reported in-significant associations (Zang et al. 2021; Zhang et al. 2021a). For example, Li et al. conducted a cross-sectional study of Chinese rural adults and found each $1 \mu\text{g}/\text{m}^3$ increase in PM_1 was associated with a 4.3% (OR = 1.403, 95%CI 1.033, 1.053) increase in hypertension (Li et al. 2019). Yang et al. conducted a cross-sectional study of 15,477 Chinese adults and found each inter-quartile range (IQR, $15 \mu\text{g}/\text{m}^3$) increase in PM_1 was associated with a 13% (OR = 1.13, 95%CI 1.04, 1.22) increase in diabetes prevalence (Yang et al. 2018a). However, Zhang et al. conducted a cross-sectional study of 9,897 Chinese children and adolescents and did not observe statistically significant associations between PM_1 and metabolic syndrome (OR = 1.20, 95%CI 0.99, 1.46) (Zhang et al. 2021a). Moreover, those studies focused on the association of long-term exposure to PM_1 with each metabolic disease (such as obesity, diabetes, hypertension, or metabolic syndrome) were published in isolation, comprehensive evidence of the potential adverse effects of PM_1 in on different metabolic diseases remains scarce, especially in rapidly increasing population of Chinese middle-aged and older adults (Niu et al. 2022a).

Therefore, to fill this gap, we used a two-stage analytic approach to investigate the associations of exposure to PM_1 with the four most common metabolic diseases among Chinese middle-aged and older adults, including obesity, diabetes, hypertension, dyslipidemia, and metabolic syndrome. Potential modification effects of health behaviors and lifestyles were also examined in this study.

Material and methods

Study population

The study population was drawn from the China Health and Retirement Longitudinal Study (CHARLS), a national cohort study of middle-aged and older adults in China. Briefly, participants were recruited from community-based populations in 28 provinces of China through a multi-stage probability proportionate sampling (PPS) method. The national baseline survey was carried out in 2011, and 3 waves of follow-ups were completed in 2013, 2015, and 2018, respectively. Details of the CHARLS are shown in previous studies (Chen et al. 2019; Zhao et al. 2014). Physical examination and blood-based biomarkers assessment were only conducted in 2011 and 2015, and estimated PM_1 concentration data was obtained after 2013; therefore, we conducted a cross-sectional study to investigate the association of PM_1 and metabolic diseases, using the health data of CHARLS wave 3 in 2015.

A total of 21,059 participants were recruited in the CHARLS 2015, with 16,406 participants participating in the physical examination. Subsequently, we excluded 2,055 study subjects without blood samples and 3,560 study subjects without metabolic biomarkers data. Finally, we excluded 351 participants in some regions of Xinjiang and Inner Mongolia where PM₁ exposure failed to be assessed. A total of 12,495 participants from 27 provinces were included in the study. The flowchart of participants' inclusion is presented in Figure S1.

Study ethics approval of the CHARLS study was obtained from Peking University (Reference Number: IRB00001052–11015). All participants and their guardians signed consent voluntarily.

PM₁ exposure assessment

Individual PM₁ exposure was estimated from space using space–time extremely randomized trees (STET) models. Briefly, surface PM₁ measurement (collected from the China Atmosphere Watch Network of the China Meteorological Administration), Multi-Angle Implementation of Atmospheric Correction aerosol optical depth (MAIAC AOD, collected from the National Aeronautics and Space Administration's (NASA) Terra and Aqua Moderate Resolution Imaging Spectroradiometer (MODIS) C6 MAIAC Level (L) 2 swath aerosol products (MCD19A2)) product, meteorological data (from ERA-Interim atmospheric reanalysis products), the multi-resolution emission inventory for China (MEIC) data, and other ancillary data (such as land cover, topography, traffic, and population) were combined in PM₁ assessments at 1 km spatial resolution. After data integration, all selected factors were input to the STET model, which took into account the spatial autocorrelation, adjacent pixels, and temporal difference. The average cross-validation R^2 and root-mean-square error (RMSE) were 0.83 and 9.50 $\mu\text{g}/\text{m}^3$ at daily levels, respectively. Detailed information on modeling methodology was shown in previous studies (Wei et al. 2019, 2023) and the CHAP website (<https://weijiang-rs.github.io/product.html>). Annual PM₁ concentrations were assigned to each participant according to their residential cities. The two-year average concentration of PM₁ was calculated as long-term PM₁ exposure (Zheng et al. 2022).

Assessment of metabolic diseases

Metabolic disease is defined as a cluster of metabolic system disorders. To identify the most common and impactful metabolic diseases in humans, we conducted a careful literature review and found that obesity, diabetes, hypertension, dyslipidemia, and non-alcoholic fatty liver disease (NAFLD) were the most notable metabolic diseases (Azizi et al. 2019; Chew et al. 2023; Steenblock et al. 2021). In addition to

NAFLD which was not available in the CHARLS dataset, our study included 4 above metabolic diseases (obesity, diabetes, hypertension, and dyslipidemia), which may effectively represent metabolic diseases. Moreover, we noticed that metabolic syndrome is defined as a cluster of metabolic disorders, including abdominal obesity, dyslipidemia, hypertension, and diabetes, has attracted wide attention from recent studies, and metabolic syndrome could reflect the metabolic disorders more comprehensively (Eckel et al. 2005; Han et al. 2022). Therefore, we included metabolic syndrome in our study.

Waist circumference (WC) was assessed for diagnosis of abdominal obesity and metabolic syndrome. Three times blood pressure measurements were completed for the diagnosis of hypertension and metabolic syndrome. Fasting blood glucose (FBG) and glycosylated hemoglobin, type A1C (HbA1C) were tested for diagnosis of diabetes and metabolic syndrome. Detailed information on WC, blood pressure, and blood-based biomarkers assessments was described in previous studies (Chen et al. 2019; Han et al. 2022; Niu et al. 2022a). In addition, self-reported clinical diseases and medication information were also collected to assist in the diagnosis of diabetes, hypertension, and metabolic syndrome via a face-to-face interview.

Abdominal obesity was defined as WC \geq 85 cm for females and WC \geq 90 cm for males (Cao et al. 2021). Diabetes cases were defined as FBG \geq 7.0 mmol/L and/or HbA1C \geq 6.5% and/or having self-reported clinically diagnosed diabetes and/or currently taking anti-diabetes medication (Association 2021; Chen et al. 2022). Hypertension was defined as SBP \geq 140 mmHg and/or DBP \geq 90 mmHg and/or having self-reported clinically diagnosed hypertension and/or currently taking anti-hypertension medication (Chew et al. 2023; Niu et al. 2022a). Dyslipidemia was defined as individual with one or more abnormal blood lipid concentrations (hypercholesterolemia: TC \geq 6.22 mmol/L; hypertriglyceridemia: TG \geq 2.26 mmol/L; hypoalphalipoproteinemia: HDL-C $<$ 1.04 mmol/L; hyperbetalipoproteinemia LDL-C \geq 4.14 mmol/L) (Mao et al. 2020b; Nie et al. 2023). Metabolic syndrome case was diagnosed as participant with WC \geq 80 cm for females and WC \geq 90 cm for males plus any two of the following components was defined as metabolic syndrome patient: (1) TG $>$ 1.7 mmol/L, (2) HDL $<$ 1.29 mmol/L in female or HDL $<$ 1.03 mmol/L in male, (3) SBP \geq 130 mmHg and/or DBP \geq 85 mmHg and/or currently taking anti-hypertension medication, and (4) FBG \geq 5.6 mmol/L and/or currently taking anti-diabetic medication (Han et al. 2022; Hou et al. 2020).

Covariates

Based on previous studies of particulate matter and metabolic diseases and directed acyclic graph analysis

(Figure S2), a series of confounders were identified in our study. These covariates included (1) climate condition: temperature and relative humidity (Han et al. 2022; Wang et al. 2023); (2) sociodemographic variables: age and sex (Liu et al. 2019b; Mao et al. 2020b; Zhang et al. 2021a); (3) socioeconomic factors: education status (Zhang et al. 2021a), marital status (Chen et al. 2022; Liu et al. 2019b), and annual household expenditure (Han et al. 2022); (4) health behaviors and lifestyles: smoking status (Yang et al. 2018b), drinking status (Mao et al. 2020b), cooking fuel use, and physical activity (PA) (Cao et al. 2021; Zhang et al. 2023). For smoking status, individuals were classified as smokers if they had ever engaged in tobacco chewing, pipe smoking, or self-rolled cigarettes, cigarettes, or cigars. Drinking status was determined by assessing participants' alcohol consumption behavior in the past year, including liquor, wine, or beer intake (Niu et al. 2022a). Physical activity was measured using the International Physical Activity Questionnaire. The physical activity score (PA score) was calculated using metabolic equivalent multipliers (MET) as follows: PA score (MET-hour/week) = $3.3 \times$ total walking weekly duration score + $4.0 \times$ total moderate activity weekly duration score + $8.0 \times$ total vigorous activity weekly duration score (Han et al. 2022; Zhang et al. 2023). For cooking fuel use, natural gas, marsh gas, and liquefied petroleum gas were classified as "clean fuel," while coal, crop residue, or wood burning were classified as "solid fuel" (Han et al. 2022; Wang et al. 2023). Covariates of temperature, relative humidity, age, and physical activity were included in models as continuous variables, while sex, education status, drinking status, and cooking fuel use were included in models as categorical variables. To avoid excessive loss of samples, missing data of covariates were input using the R package "mice" in this study (Wang et al. 2023).

Statistical analysis

Descriptive statistics of participants were conducted. Continuous variables were described as mean \pm standard deviation (SD), and categorical variables were described as frequency (percentage, %). Student's *t*-test was used to examine the difference of continuous variables between patients and non-patients, and the Chi-square test was applied to assess the distribution discrepancy of categorical variables. Mean, SD, and interquartile range (IQR) were applied to summarize the descriptive characteristics of metabolic biomarkers, PM₁ exposure, and climate conditions.

A two-stage analytic approach was performed to explore the relationship between PM₁ exposure and metabolic diseases. In the first stage, generalized linear models were performed to examine the associations of PM₁ concentrations with metabolic disease risks, including abdominal obesity, diabetes, hypertension, and metabolic syndrome.

Results were shown as the odds ratio (OR) and its 95% confidence interval (CI) of metabolic disease risk with each 10 $\mu\text{g}/\text{m}^3$ increase in PM₁ concentrations. First, a crude model was performed without any adjustment. Second, temperature and relative humidity were included in adjusted model 1 to adjust potential climate condition confounders using the spline function. Third, sociodemographic and socioeconomic factors were additionally adjusted in adjusted model 2. Finally, health behaviors and lifestyle variables were additionally included in adjusted model 3. In the second stage, concentration–response (C-R) relationships of long-term exposure to PM₁ with the OR of metabolic diseases were also investigated in this study. Specifically, relationships of PM₁ with metabolic diseases were fitted as a smoothing term using restricted cubic spline (RCS) with 3 knots at the 10th, 50th, and 90th percentiles of 2-year average PM₁ concentration (reference is the 10th percentile) (Wang et al. 2023). Nonlinearity was tested using analysis of variance (ANOVA), and inflection points were identified using segmented regression (Hou et al. 2023).

Previous studies indicated that health lifestyles, such as smoking (Mao et al. 2020b), alcohol consumption (Mao et al. 2020b), cooking fuel use (Cao et al. 2021; Liu et al. 2021), and physical exercise (Hou et al. 2020; Zhang et al. 2023), could modify the adverse of air pollution on metabolic diseases; we conducted a series of subgroup analysis by different health lifestyles (smoking status, drinking status, cooking fuel use, and PA) to provide more evidence for reducing the adverse effect of exposure to PM₁ on metabolic diseases by choosing more healthy lifestyles.

Several sensitivity analyses were conducted to test the robustness of the results. First, we applied a generalized linear mixed-effect model to re-examine the association between PM₁ and metabolic diseases by including City ID as a random-effect intercept. This model allows participants in the same city to serve as their own control and could adjust for the spatial clustering of metabolic diseases (Ren et al. 2023; Wu et al. 2020). Second, we re-investigated the associations of PM₁ with metabolic diseases by using 1-year average concentration of PM₁ exposure before CHARLS Wave 3. Third, log-binomial Poisson regressions were performed, accounting for the high prevalence of metabolic disease (Barros and Hirakata 2003, Han et al. 2022). Fourth, participants who had ever changed their address from their last investigation in 2013 were excluded. Fifth, associations of PM₁ with hypertension and metabolic syndrome were re-examined by excluding participants currently taking anti-hypertension medication, accounting for the potential influence of anti-hypertension medication on blood pressure (Niu et al. 2022a). Finally, the associations of PM₁ with diabetes and metabolic syndrome were assessed by excluding participants currently taking anti-diabetic medication, accounting

for the potential influence of anti-diabetic medication on blood glucose.

Statistical analyses of this study were completed using R (version 4.2.1). “mice” package for the imputation of missing data, and “plotRCS” for C-R relationships plots. A two-sided *P* value < 0.05 was regarded as statistical significance.

Results

Descriptive statistic of study population characteristics, metabolic biomarkers, and ambient PM₁ exposure

A total of 12,495 participants were included in this study, and these participants were recruited from 435 communities in 123 cities. The spatial distribution of participants in 27 provinces of China is presented in Fig. 1. Of the 12,495 participants, 5,714 (45.73%) adults were identified as abdominal obesity cases, 2,526 (20.22%) adults were identified as diabetes cases, 5,303 (42.46%) adults were identified as hypertension cases, 5,124 (41.01%) adults were identified as dyslipidemia cases, and 4,221 (33.78%) adults were identified as metabolic syndrome cases, respectively. The mean age of participants was 58.79 ± 13.14 years. Table 1 summarizes the descriptive characteristics of the participants.

Table 2 presents the descriptive characteristics of metabolic biomarkers, PM₁ exposure, and climate conditions. In terms of metabolic biomarkers, the mean WC, FBG, HbA1c, SBP, DBP, TG, TC, HDL, and LDL were 85.24 ± 13.16 cm, 5.75 ± 1.95 mmol/L, 5.87 ± 1.00%, 128.33 ± 19.66 mmHg, 75.43 ± 11.19 mmHg, 1.61 ± 1.03 mmol/L, 4.75 ± 0.94 mmol/L,

1.32 ± 0.29 mmol/L, and 2.64 ± 0.74 mmol/L, respectively. The 2-year average PM₁ concentration of participants was 37.51 ± 10.04 µg/m³.

Associations between exposure to PM₁ and metabolic diseases

Figure 2 presents the associations between PM₁ and four metabolic diseases. Both crude models and three adjusted models indicated that exposure to higher PM₁ concentrations was associated with increased risks of abdominal obesity, diabetes, hypertension, and metabolic syndrome. In fully-adjusted model (adjusted model 3), each 10 µg/m³ increase in PM₁ concentrations was associated with 39% (OR = 1.39, 95% CI 1.33, 1.46) increase in abdominal obesity, 18% (OR = 1.18, 95%CI 1.12, 1.25) increase in diabetes, 11% (OR = 1.11, 95%CI 1.06, 1.16) increase in hypertension, and 25% (OR = 1.25, 95%CI 1.19, 1.31) in metabolic syndrome, respectively.

C-R relationships of exposure to PM₁ with metabolic diseases

Figure 3 presents the C-R relationships of exposure to PM₁ with metabolic diseases. We observed that the OR values of abdominal obesity, diabetes, hypertension, and metabolic syndrome were increased gradually with the increase of PM₁ concentrations (*P* for overall < 0.05), while a similar trend was not been observed in the relationship of PM₁ with dyslipidemia (*P* for overall > 0.05). In terms of the shapes of C-R curves, the C-R curves of PM₁ exposure with abdominal obesity, diabetes, and metabolic syndrome were nearly

Fig. 1 The spatial distribution of 12,495 middle-aged and older adults in 27 provinces of China

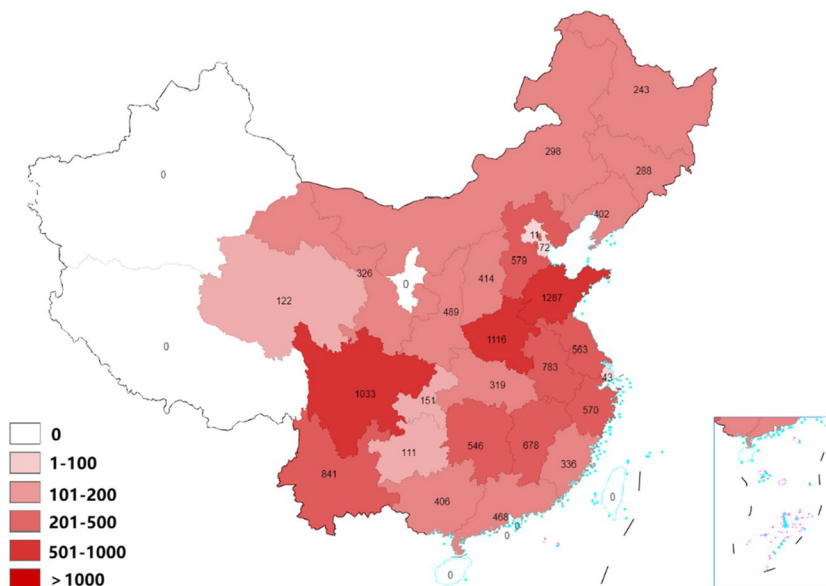


Table 1 Basic characteristics of study participants

Characteristics	Abdominal obesity		Diabetes		Hypertension		Dyslipidemia		Metabolic syndrome	
	Yes (n = 5714)	No (n = 6781)	Yes (n = 2526)	No (n = 9969)	Yes (n = 5303)	No (n = 7192)	Yes (n = 5124)	No (n = 7371)	Yes (n = 4221)	No (n = 8274)
Age (mean ± SD)										
Age (years)	58.79 ± 13.14	58.42 ± 12.68	59.10 ± 13.50*	58.30 ± 12.98***	61.46 ± 13.61	56.81 ± 12.41***	59.65 ± 12.48	58.73 ± 13.18*	58.94 ± 13.22	58.70 ± 13.01
Sex (n (%))										
Male	5732 (45.9)	2032 (35.6)	3700 (54.6)***	4614 (46.3)	2511 (47.4)	3221 (44.8)**	2545 (49.7)	3187 (43.2)***	1313 (31.1)	4419 (53.4)***
Female	6763 (54.1)	3682 (64.4)	3081 (45.4)	5355 (53.7)	2792 (52.6)	3971 (55.2)	2579 (50.3)	4184 (56.8)	2908 (68.9)	3855 (46.6)
Residence (n (%))										
Rural	7760 (62.1)	3199 (56.0)	4561 (67.3)***	6344 (63.6)***	3189 (60.1)	4571 (63.6)***	3112 (60.7)	4638 (62.9)**	2290 (54.3)	5470 (66.1)***
Urban	4735 (37.9)	2515 (44.0)	2220 (32.7)	3625 (36.4)	2114 (39.9)	2621 (36.4)	2012 (39.3)	2732 (37.1)	1931 (45.7)	2804 (33.9)
Education status (n (%))										
Elementary or below	8267 (66.2)	3704 (64.8)	4563 (67.3)*	6544 (65.6)	3670 (69.2)	4597 (63.9)***	3361 (65.6)	4906 (66.6)	2842 (67.3)	5425 (65.6)*
Middle school or high	4228 (33.8)	2010 (35.2)	2218 (32.7)	3425 (34.4)	1633 (30.8)	2595 (36.1)	1763 (34.4)	2465 (33.4)	1379 (32.7)	2849 (34.4)
Marital status (n (%))										
Married	10,863 (86.9)	4986 (87.3)	5877 (86.7)	8709 (87.4)**	4420 (83.3)	6443 (89.6)***	4443 (86.7)	6420 (87.1)	3632 (86.0)	7231 (84.7)*
Single, divorced, and widowed	1632 (13.1)	728 (12.7)	904 (13.3)	1260 (12.6)	883 (16.7)	749 (10.4)	681 (13.3)	951 (12.9)	589 (14.0)	1043 (12.6)
Household expenditure (mean ± SD)										
Household expenditure (×10 ³ CNY)	14.22 ± 41.40	13.92 ± 35.11	14.48 ± 46.05	14.22 ± 43.07	12.80 ± 32.63	15.27 ± 46.81**	14.82 ± 47.70	13.81 ± 36.39	14.12 ± 37.25	14.28 ± 43.38
Smoking status (n (%))										
Non-smoker	7234 (57.9)	3793 (66.4)	3441 (50.7)***	5773 (57.9)	2973 (56.1)	4261 (59.2)***	2766 (54.0)	4468 (60.6)***	2916 (69.1)	4318 (52.2)***
Smoker	5261 (42.1)	1921 (33.6)	3340 (49.3)	4196 (42.1)	2330 (43.9)	2931 (408)	2358 (46.0)	2903 (39.4)	1305 (30.9)	3956 (47.8)
Drinking status (n (%))										
Non-drinker	3262 (26.0)	1275 (22.3)	1977 (29.2)***	2674 (26.8)***	1388 (26.2)	1864 (25.9)	1331 (26.0)	1921 (26.1)	830 (19.7)	2422 (29.3)***
Drinker	9243 (74.0)	4433 (77.6)	4800 (70.8)	7286 (73.1)	3909 (73.7)	5324 (74.0)	3789 (73.9)	5444 (73.9)	3388 (80.3)	5845 (70.6)
Cooking fuel use (n (%))										
Solid fuel	2700 (25.8)	3496 (61.2)	3822 (56.4)***	5791 (58.1)*	3076 (58.0)	4242 (59.0)	3025 (59.0)	4293 (58.2)	2615 (62.0)	4703 (56.8)***
Clean fuel	3672 (35.1)	2218 (38.8)	2959 (43.6)	4178 (41.9)	2227 (42.0)	2950 (41.0)	2099 (41.0)	3078 (41.8)	1606 (38.0)	3571 (43.2)
Physical activity (mean ± SD)										
PA score (MET-hour/week)	125.89 ± 109.20	117.68 ± 105.30	132.82 ± 111.60***	127.64 ± 110.21***	121.39 ± 107.58	129.22 ± 109.96*	117.67 ± 107.33	126.44 ± 109.12*	116.76 ± 103.97	130.56 ± 111.23***

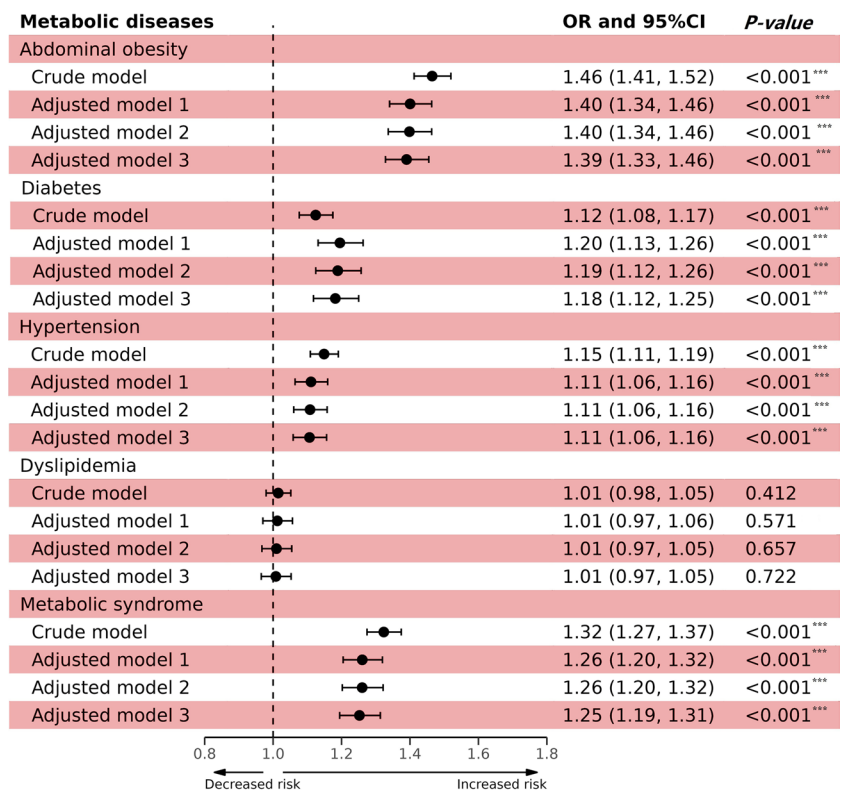
***, ***, and * represented the P value of difference comparison for the characteristics between patients and non-patients; * P < 0.05, ** P < 0.01, and *** P < 0.001

Table 2 Descriptive statistics of metabolic biomarkers, PM₁, and meteorological factors

Variables	Mean ± SD	P25	P50	P75	IQR
Metabolic bio-markers					
WC (cm)	85.24 ± 13.16	78.80	86.00	93.00	14.20
FBG (mg/dL)	103.53 ± 35.04	88.29	95.50	106.30	18.01
SBP (mmHg)	128.33 ± 19.66	114.00	126.00	140.33	26.33
DBP (mmHg)	75.43 ± 11.19	67.33	74.67	82.33	15.00
TG (mmol/L)	1.61 ± 1.03	0.93	1.29	1.92	0.99
TC (mmol/L)	4.75 ± 0.94	4.10	4.67	5.30	1.20
HDL (mmol/L)	1.32 ± 0.30	1.12	1.29	1.49	0.37
LDL (mmol/L)	2.64 ± 0.74	2.13	2.59	3.09	0.96
Air pollutant					
PM ₁ (µg/m ³)	37.51 ± 10.04	30.48	37.37	44.66	14.18
Climate condition					
Temperature (°C)	15.43 ± 4.03	14.10	16.10	17.85	3.75
Relative humidity (%)	67.84 ± 10.89	60.50	70.50	76.00	15.50

linear (*P* for nonlinear > 0.05). The C-R curves of PM₁ exposure with hypertension showed slight “J” shapes (*P* for nonlinear = 0.010), with inflection points of 37.5 µg/m³.

Fig. 2 OR and 95%CI of metabolic diseases associated with 10 µg/m³ increase in PM₁. Notes: **P* < 0.05, ***P* < 0.01, and ****P* < 0.001



Subgroup analysis by different health behaviors and lifestyles for the associations of PM₁ concentration with metabolic diseases

Figure 4 displays the subgroup analysis by different health behaviors and lifestyles for the associations of PM₁ with metabolic diseases. Positive and significant associations of PM₁ with abdominal obesity, diabetes, hypertension, and metabolic syndrome were observed in all subgroups. When comparing the estimated effects in different subgroups, we observed the OR value of metabolic syndrome with PM₁ was higher in solid fuel use participants than that of clean fuel use participants, with *P*-interaction < 0.05.

Sensitivity analysis

Sensitivity analysis showed that the positive associations of PM₁ with all four metabolic diseases were stable when City ID was included in a generalized linear mixed-effect model as a random-effect intercept, 1-year average PM₁ concentration, and log-binomial Poisson regression model were applied in models or participants who had ever changed their address from the last investigation were excluded (Table S2-S5). Sensitivity analysis by excluding anti-hypertension medication or anti-diabetic medication takers showed consistent associations, except for hypertension in adjusted model 1 and diabetes in the crude model (Table S6, Table S7).

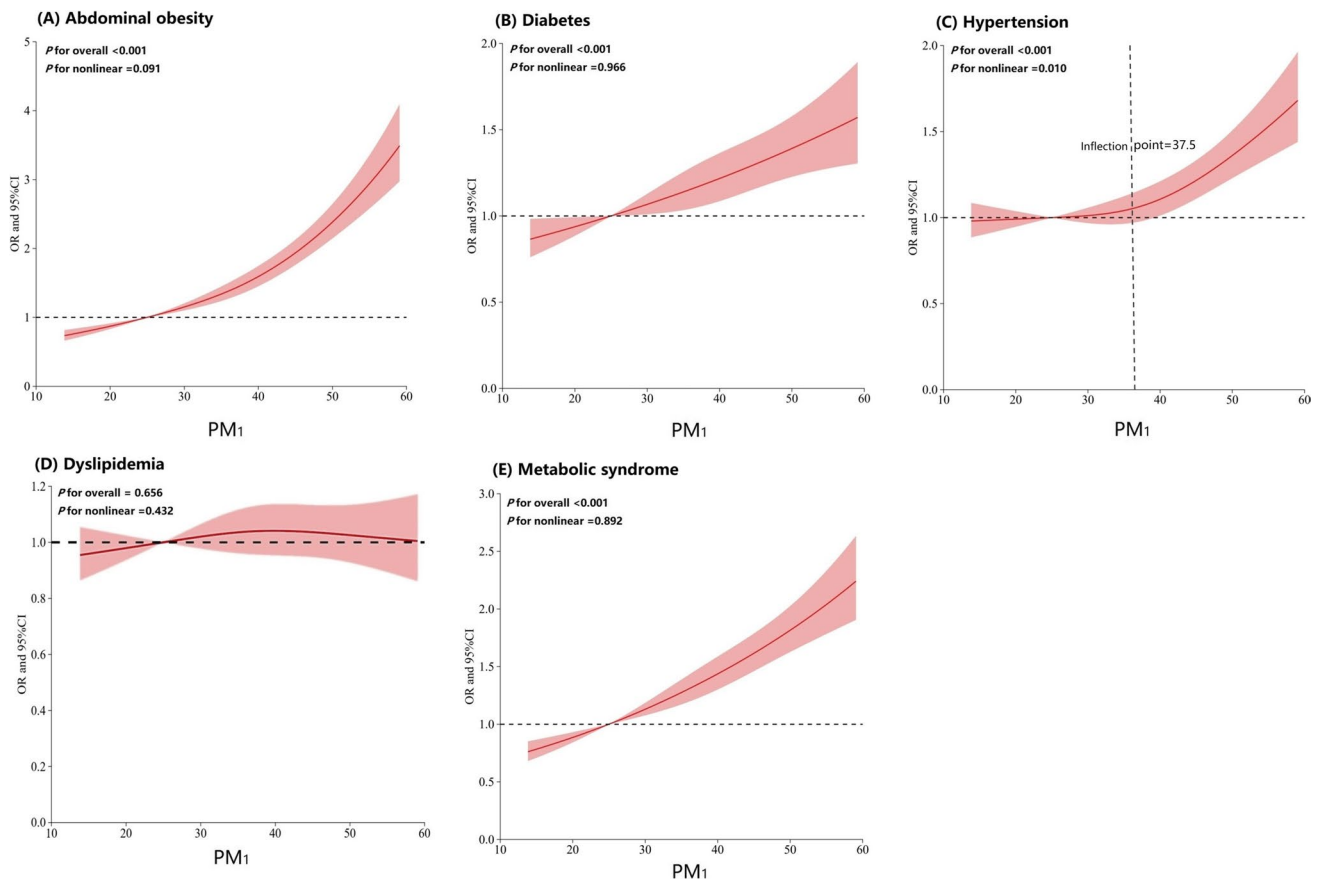


Fig. 3 Concentration–response relationships of long-term exposure to PM_{10} with the OR of metabolic diseases. **A** abdominal obesity, **B** diabetes, **C** hypertension, **D** dyslipidemia, and **E** metabolic syndrome

Discussion

Key findings

In this Chinese national cross-sectional study, we found that exposure to higher PM_{10} concentration was associated with an increased risk of metabolic diseases, including abdominal obesity, diabetes, hypertension, and metabolic syndrome. Moreover, the C-R curves of PM_{10} with metabolic diseases showed that the OR values of metabolic diseases were increased gradually with the increase of PM_{10} concentrations, with slight “J” shapes for hypertension and linear shapes for abdominal obesity, diabetes, and metabolic syndrome. To our knowledge, this is the most comprehensive study that has systematically examined the adverse effects of PM_{10} on four common metabolic diseases.

Comparison with other studies and interpretations

While the associations of exposure to PM_{10} with different metabolic diseases still have not been examined in one single

study, several studies of long-term exposure to PM_{10} with specific metabolic diseases could support our findings.

We found that exposure to higher PM_{10} concentration was associated with an increased abdominal obesity risk. This finding was consistent with three similar designed studies (Liu et al. 2023; Zhang et al. 2021a, 2021b). In a national Chinese cross-sectional study, Liu et al. reported that each $10 \mu\text{g}/\text{m}^3$ increase in PM_{10} concentrations was associated with a 29.4% (OR = 1.294, 95%CI 1.225, 1.367) increase in the prevalence of central obesity (Liu et al. 2023). Moreover, two cross-sectional studies of children and adolescents also reported a positive association between PM_{10} and abdominal obesity. They found that each $10 \mu\text{g}/\text{m}^3$ increase in PM_{10} concentrations was associated with a 42% (OR = 1.42, 95%CI 1.23, 1.64) (Zhang et al. 2021a) and 39% (OR = 1.39, 95%CI 1.23, 1.57) (Zhang et al. 2021b) increase in the prevalence of central obesity, which were similar to our findings (OR = 1.39, 95%CI 1.33, 1.46). The estimated effects of our study were similar to the studies conducted by Zhang et al. (2021a) and Zhang et al. (2021b), but higher than that of Liu et al.’s study, which might be due to different definition of obesity, PM_{10} exposure estimation methods, different sample

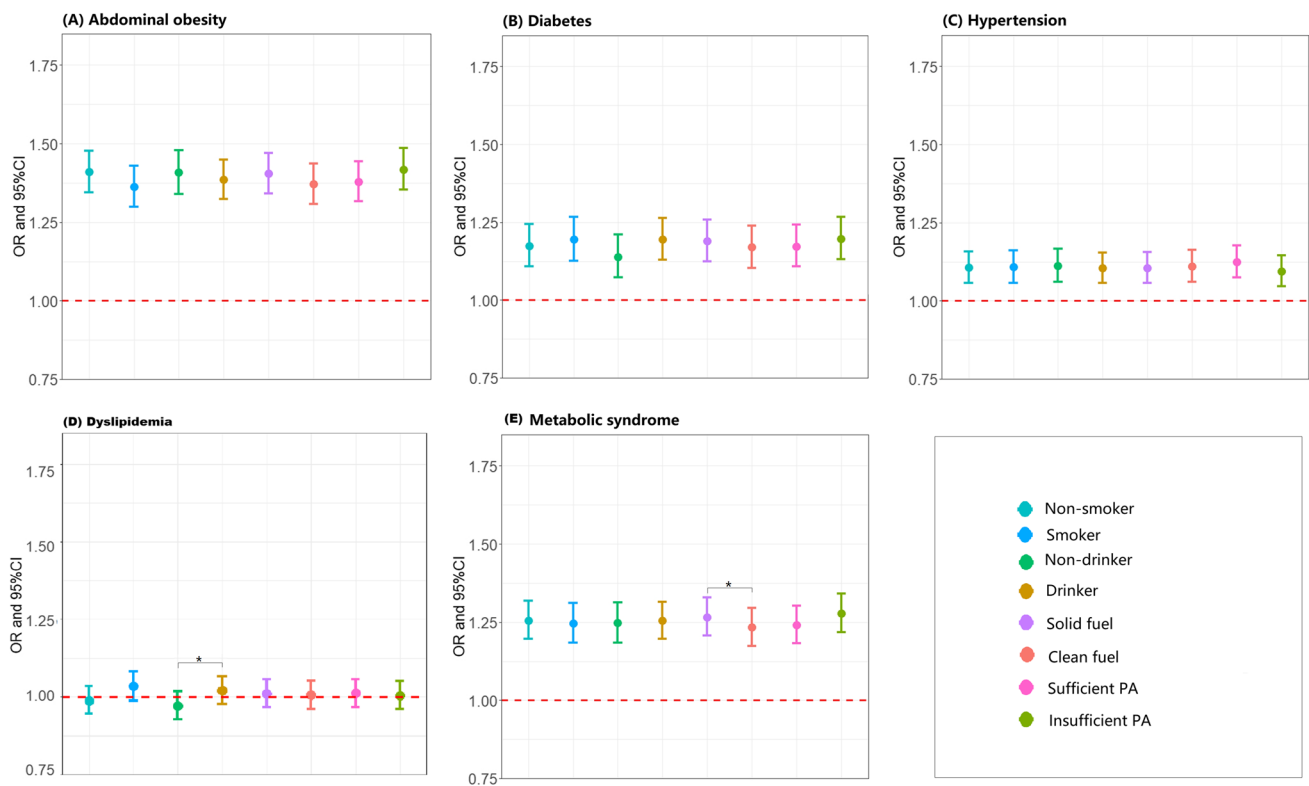


Fig. 4 Subgroup analysis by lifestyles for the associations of long-term exposure to PM_{10} with metabolic diseases. **A** abdominal obesity, **B** diabetes, **C** hypertension, **D** dyslipidemia, and **E** metabolic syndrome. Notes: * $P < 0.05$, ** $P < 0.01$, and *** $P < 0.001$

size, and other study characteristics. In addition, we found that those studies were conducted in China; further studies are necessary to confirm the adverse effects of exposure to PM_{10} on metabolic systems in other countries, especially low-polluted countries. Finally, we also examined the C-R of PM_{10} with abdominal obesity and observed linear shapes of C-R curves for the relationship of PM_{10} with abdominal obesity. To our knowledge, the C-R of PM_{10} with abdominal obesity has not been reported previously, and our study might provide more comprehensive evidence for the adverse effects of long-term exposure to PM_{10} on abdominal obesity risk.

This study indicated that long-term exposure to PM_{10} was related to increased diabetes risk. Several previous studies have examined the association of exposure to PM_{10} with diabetes, with some studies finding positive associations (Liu et al. 2019b; Liu et al. 2019c; Mei et al. 2023; Yang et al. 2018a) and other studies reporting nonsignificant association (Wang et al. 2022). For example, Mei et al. conducted a cross-sectional study and found that each $10 \mu\text{g}/\text{m}^3$ increase in PM_{10} was associated with a 14% (OR = 1.14, 95%CI 1.0, 1.29) increase in diabetes risk (Mei et al. 2023). Another cross-sectional study of the 33 Chinese communities indicated that each $15 \mu\text{g}/\text{m}^3$ increase in PM_{10} was associated with a 13% (OR = 1.13, 95%CI 1.04, 1.29) increase in

diabetes risk (Yang et al. 2018a). The estimated effects of our study (OR = 1.18, 95%CI 1.12, 1.25), but higher than that of Mei et al.'s study and Yang et al.'s study, which may be explained by different study sample size, study period, and regression models. Conversely, a cross-sectional study from the China Multi-Ethnic Cohort reported an insignificant association between PM_{10} and diabetes (Wang et al. 2022). The insignificant result of Wang et al.'s study may be explained by sample size, study population, and region. In addition, we examined the C-R of PM_{10} with diabetes. It is worth noting that the C-R curve was nearly linear, suggesting that even exposure to low-level PM_{10} might lead to increased diabetes risk (Meng et al. 2021).

Positive relationships between PM_{10} and hypertension risk identified in this study were reported in previous studies (Li et al. 2019; Niu et al. 2023; Yang et al. 2019). For example, a cross-sectional study of Chinese rural adults found that each $1 \mu\text{g}/\text{m}^3$ increase in PM_{10} concentration was associated with a 4.3% (OR = 1.043, 95%CI 1.033, 1.053) increase in hypertension (Li et al. 2019). Two meta-analyses also examined the association of long-term exposure to PM_{10} with the prevalence of hypertension (Niu et al. 2022b; Qin et al. 2021). Qin et al. conducted a meta-analysis of 2 studies and found no significant association of PM_{10} with hypertension (Qin et al. 2021). The insignificant result of Qin et al.'s study

may be due to the limited original studies. After Qin's study, we also conducted a meta-analysis of 4 studies and found a positive association between PM_{10} with hypertension (Niu et al. 2023). This meta-analysis showed that each $10 \mu\text{g}/\text{m}^3$ increase in PM_{10} was associated with a 27% (OR = 1.27, 95%CI 1.06, 1.52) increase in hypertension risk, which was significantly higher than that of this study (OR = 1.11, 95%CI 1.06, 1.16). The inconsistency of the estimated effect might be due to the limited number of included studies (only 4 studies) and significant heterogeneity among the included studies (Niu et al. 2023). Moreover, our study also observed a slight "J" shape C-R curve for the relationships of PM_{10} with hypertension. While no study examined the C-R for the relationships between PM_{10} with hypertension, our findings could be supported by Song et al.'s study (Song et al. 2021). They found the C-R functions between $PM_{2.5}$ concentration and hypertension risk were presented as a slight "J" shape, which was consistent with our findings.

Several previous studies have examined the association of exposure to PM_{10} with dyslipidemia (Gui et al. 2020; Hu et al. 2023; Mao et al. 2020b; Sun et al. 2023; Wang et al. 2021). For example, a cross-sectional study of 67,305 adults in the China Multi-Ethnic Cohort (CMEC) study found that exposure to PM_{10} was associated with increased dyslipidemia risk. They found that each $10 \mu\text{g}/\text{m}^3$ increase in PM_{10} was associated with a 13% (OR = 1.13, 95%CI 1.06, 1.21) increase in dyslipidemia risk (Wang et al. 2021). A longitudinal study of 6976 middle-aged and older adults found that exposure to PM_{10} was associated with increased dyslipidemia incidence. They found that each $10 \mu\text{g}/\text{m}^3$ increase in PM_{10} was associated with a 3% (OR = 1.03, 95%CI 1.01, 1.23) increase in dyslipidemia incidence (Hu et al. 2023). However, a cross-sectional study of Chinese rural adults reported that exposure to PM_{10} was associated with decreased dyslipidemia (Mao et al. 2020b). In this study, we did not observe a statistically significant association between exposure to PM_{10} and dyslipidemia risk. The inconsistency of results may be explained by various causes, such as diversity in the samples, study geographies, and methodology. Moreover, as a comprehensive indicator of blood lipid disorder, dyslipidemia is defined as an individual with one or more abnormal blood lipid concentrations, including hypercholesterolemia, hypertriglyceridemia, hypoalphalipoproteinemia, and hyperbetalipoproteinemia. The inconsistency of association between PM_{10} and dyslipidemia may be attributed to the different results of these four indicators. For example, a cross-sectional study of 12,814 children and adolescents found exposure to PM_{10} was associated with increased hypercholesterolemia risk (OR = 2.15, 95%CI 1.27, 3.65), but the associations of PM_{10} with hypertriglyceridemia, hypoalphalipoproteinemia, and hyperbetalipoproteinemia were non-significant (Gui et al. 2020). Similar inconsistent results of PM_{10} with these four indicators were also reported by Mao

et al. (Mao et al. 2020b). Finally, as a recent meta-analysis mentioned, although PM_{10} was associated with increased dyslipidemia risk, PM_{10} was weakly correlated with four indicators of dyslipidemia (Sun et al. 2023). Further studies with large-scale and more regions should be conducted to provide more convincing evidence on the associations between PM_{10} and dyslipidemia risk.

Five previous studies have investigated the associations of PM_{10} with metabolic syndrome (Hou et al. 2020; Liu et al. 2023; Yang et al. 2018b; Zang et al. 2021; Zhang et al. 2021a). However, the results are still controversial. Three studies of Chinese adults reported positive associations between PM_{10} and metabolic syndrome risk (Hou et al. 2020; Liu et al. 2023; Yang et al. 2018b). For example, Hou et al. conducted a cross-sectional study of 39,089 rural adults and indicated that each $5 \mu\text{g}/\text{m}^3$ increase in PM_{10} concentration was associated with a 25.1% (OR = 1.251, 95%CI 1.199, 1.306) increase in metabolic syndrome (Hou et al. 2020). In our study, we found that each $10 \mu\text{g}/\text{m}^3$ increase in PM_{10} was associated with a 25% (OR = 1.25, 95%CI 1.19, 1.31) increase in metabolic syndrome. We found that the estimated effect of our study was lower than that of Hou et al.'s study, which may be explained by more serious PM_{10} pollution (the average PM_{10} concentration $57.45 \mu\text{g}/\text{m}^3$ for Hou et al.'s study, and $37.51 \mu\text{g}/\text{m}^3$ for our study). However, significant associations between PM_{10} and the prevalence of metabolic syndrome have not been found in a meta-analysis (Zang et al. 2021), children, and adolescents (Zhang et al. 2021a). The in-significant result of Zang et al.'s study may be explained by only 2 original research included, and the in-significant result of Zhang et al.'s study may be due to the different age composition of participants, different diagnostic criteria of metabolic syndrome in participants of different age groups. In our study, we found that each $10 \mu\text{g}/\text{m}^3$ increase in PM_{10} was associated with a 25% (OR = 1.25, 95%CI 1.19, 1.31) increase in metabolic syndrome, and the C-R curve of PM_{10} with metabolic syndrome risk was nearly linear, which might provide new epidemiological evidence of adverse effects of long-term exposure to PM_{10} on metabolic syndrome risk.

In addition to the robust positive associations of PM_{10} exposure with abdominal obesity, diabetes, hypertension, and metabolic syndrome in all subgroups, we also indicated that cooking fuel use can modify the adverse effects of PM_{10} on metabolic syndrome. Specifically, for each $10 \mu\text{g}/\text{m}^3$ increase in PM_{10} , the OR values of metabolic syndrome were 1.26 (95%CI 1.21, 1.33) in solid fuel use participants and 1.23 (95%CI 1.17, 1.29) in clean fuel use participants, with P -interaction < 0.05 . Several potential explanations should be noted. First, incomplete combustion of solid fuels also could produce PM_{10} (Clark et al. 2013; Li et al. 2021), resulting in higher actual PM_{10} exposure concentrations than the estimated exposure concentrations. Second, in addition to

PM₁, solid fuel burning can release carbon monoxide, sulfur dioxide, and other air pollutants (Gordon et al. 2014); those air pollutants would also lead to increased metabolic syndrome risks. Third, solid fuel users tend to have lower socioeconomic levels and worse living conditions than clean fuel users (Gordon et al. 2014; Li et al. 2021), which would accelerate the adverse effects of PM₁ on metabolic syndrome.

Potential biological mechanisms

As PM₁ is a major constituent of PM (Mao et al. 2020b), the biological mechanisms of adverse effects of PM₁ on metabolic diseases might be interpreted by several studies of PM and metabolic diseases. Although the specific biological mechanisms underlying the impact of PM exposure on the human metabolic system have not been fully understood, several biological mechanisms, including oxidative stress (Hu et al. 2021; Rao et al. 2018), systematic inflammation (Hahad et al. 2020), dysfunction autonomic nerve system, and DNA damage, have been proposed in previous studies. First, the PM could be inhaled and deposited in the respiratory system, as well as entry into the circulatory system directly (Chen et al. 2015; Xu et al. 2022), which could trigger oxidative stress and produce a large number of reactive oxygen species (ROS). Excessive reactive oxygen species (ROS) could primarily alter the structure and function of biomacromolecules, such as lipids and proteins, leading to metabolic system disorder (Qiu et al. 2019; Wang and Tang 2019). In addition to numerous epidemiological studies (Hu et al. 2021), previous animal studies also have linked PM exposure to oxidative stress and metabolic system disorder (Piao et al. 2018; Ren et al. 2020). For example, Ren et al. conducted an experimental study in zebrafish embryos and found that the extractable organic matter from PM_{2.5} induced excessive ROS production and resulted in DNA damage and apoptosis, leading to cardiovascular system injury and metabolic system disorders (Ren et al. 2020). Second, systematic inflammation is an important biological mechanism of metabolic system disorders induced by PM (Tang et al. 2020; Xie et al. 2023). The inhaled PM could stimulate abnormal expression of inflammatory factors within cells, and the excessive release of these factors is a major cause of inflammation. Macrophages, which are an important defense barrier in the organism, could produce interleukin-1 (IL-1), interleukin-6 (IL-6), and tumor necrosis factor-alpha (TNF- α), and these factors could participate in the inflammatory response, ultimately leading to metabolic system disorders (Tang et al. 2020; Xie et al. 2023). Third, inhalation of PM may lead to the dysfunction autonomic nervous system, which has been reported as the crucial pathway of PM-induced inflammatory responses and metabolic system disorder increase (Niu et al. 2020; Xie et al. 2023).

Xie et al. conducted an animal study of diesel exhaust PM_{2.5} with pulmonary and systemic inflammation and indicated that the autonomic nervous system may play an important role in inflammatory responses induced by PM_{2.5} (Xie et al. 2023). Fourth, particulate matter-induced DNA methylation is indicated as an important physiological mechanism of the adverse effects of PM exposure on metabolic systems and diseases (Chen et al. 2016; Eze et al. 2020; Wang et al. 2023).

Strengths and limitations

To our knowledge, this is the first study that examined the associations of PM₁ with four common metabolic diseases, including abdominal obesity, diabetes, hypertension, and metabolic syndrome. This study provided comprehensive epidemiological evidence for the adverse effects of exposure to PM₁ on metabolic diseases, especially abdominal obesity and metabolic syndrome. Moreover, C-R relationships of PM₁ with metabolic diseases were examined in this study, the slight “J” shapes for hypertension, and linear shapes for abdominal obesity, diabetes, and metabolic syndrome expand current knowledge on the relationships of exposure to PM₁ with metabolic diseases. The large sample size, nationwide coverage, and high spatiotemporal resolution of PM₁ assessment provided convincing epidemiological evidence for the adverse effects of PM₁ on metabolic systems.

Several limitations should be acknowledged in the study. First, the primary limitation was the cross-sectional study design. Cause and effect of the relationship between PM₁ and metabolic diseases could not be established. Longitudinal cohort studies are warranted to establish causal inference between PM₁ exposure and metabolic disease incidence. Second, the basic characteristics of participants, history of clinically diagnosed diseases, and medication information were collected using self-reported questionnaires, thus reporting bias and recall bias might exist (Han et al. 2022). Third, diet and other health lifestyles were not available in CHARLS, which might induce metabolic biomarkers changes (Wang et al. 2023); however, smoking status, drinking status, cooking fuel use, and physical activity were adjusted in the study and results suggested that the association of PM₁ with metabolic diseases were all robust. Fourth, anti-hypertension or anti-diabetic medication intake would also influence blood pressure and blood glucose (Chew et al. 2023; Niu et al. 2022a). However, the sensitivity analysis of excluding anti-hypertension or anti-diabetic medication users showed consistent and robust results. Fifth, several participants changed their living cities, which might lead to erroneous estimates in PM₁ exposure assessment (Yao et al. 2022). However, we conducted a sensitivity analysis excluding 498 participants who changed their address after 2013 and still observed the positive associations of PM₁ with

all four metabolic diseases. Finally, dietary factors and other lifestyle factors were not collected in the CHARLS, which might modify the associations of PM₁ with metabolic diseases. Further studies that include control for more potential confounders are required to confirm our results.

Conclusion

In summary, this national cross-sectional study provides robust evidence for the adverse effects of exposure to higher PM₁ on metabolic disease risks, including abdominal obesity, diabetes, hypertension, and metabolic syndrome. Further longitudinal cohort studies are warranted to establish a causal inference between PM₁ exposure and metabolic disease risk.

Supplementary Information The online version contains supplementary material available at <https://doi.org/10.1007/s11356-023-31098-z>.

Author contribution Z.N. and Q.Z. contributed to the study conception and design; X.L., J.Z., Z.D., S.M., and J.W. contributed to the data collection; Q.Z., X.L., and J.Z. performed the data analysis and drafted the manuscript. Z.N. and S.H. helped revise the manuscript. All authors read and approved the final manuscript.

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Data availability Data will be made available on request.

Declarations

Ethical approval Study ethics approval of the CHARLS study was obtained from Peking University (Reference Number: IRB00001052–11015).

Consent to participate All participants and their guardians signed consent voluntarily.

Consent for publication Not applicable.

Competing interests The authors declare no competing interests.

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